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Research Article

Heavy Metals in Rock Oyster Saccostrea cucullata Collected from Sungai Tapai and Pantai Lido, Peninsular Malaysia: An Insight from Health Risk Assessment

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Introduction

The rock oyster, *Saccostrea cucullata* is a filter-feeder bivalve (Tack and Polk, 1996) that lives on trunks and roots of mangroves and rocky substrate (Tack *et al.*, 1992). In general, oysters are the accumulators of heavy metals (Yap et al., 2011a) and hyper accumulators for Zn (Silva et al., 2006; Yap et al., 2011b; Wang and Lu, 2017).

Numerous studies on heavy metals in the oysters are reported in the literature. In Malaysia, bioaccumulation of heavy metals in oysters was reported by Najiah *et al.* (2008) and Lim *et al.* (1995, 1998) on *Crassaostrea iredalei*; Saed *et al.* (2004) on *Isognomon alatus* and Lim *et al.* (1995,1998) on *C. belcheri.* Studies on the uses of oysters as biomonitors of heavy metal contamination were reported from Natal Brazil (Silva *et al.*, 2006), and US coast (O'connor and Lauenstein, 2006).

The distributions of heavy metals in the different organs or tissues of intertidal molluscs have been well reported in the literature. These published studies included the green-lipped mussel *Perna viridis* (Yap *et al.*, 2012, 2006; Yap, 2018), clam *Polymesoda erosa* (Yap *et al.*, 2014a), mangrove snail *Nerita lineata* (Yap and Cheng, 2013; Yap *et al.*, 2014), cockle *Anadara granosa* (Yap and Lo, 2013), and mudflat snail *Telescopium* (Yap *et al.*, 2013) and rock oyster *S. cucullata* (Yap *et al.*, 2010).

However, all these citations never discussed the human health risk assessment (HHRA) using Target Hazard Quotient (THQ) in the specific organ or tissues investigated. All of the above interpretations of results were focused on metal bio availabilities and bio monitoring points of view (Rainbow, 1995). The objective of the present study is to determine the HHRA of Cd, Cu, Fe, Ni, Pb and Zn in the edible soft tissues of *S. cucullata* collected from Sg. Tapai and Pantai Lido in 2008.



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Materials and Methods

About 25-30 of the oysters, *S. cucullata* from Sg. Tapai (Kelantan; sampled on 13 May 2008) and Pantai Lido (near Danga Bay, Johor; sampled on 3 May 2008) (Figure 1) was used for the metal analysis. The sampling sites description and some *in-situ* surface water parameters are given in (Table 1).

The identification of the oyster species was based on the book authored by Takashi (2000), and the Malaysia Fisheries Directory (2005) by Department of Fisheries Malaysia. The shell lengths (cm) of the oysters ranged from 2.82-5.54 and 6.92-9.49 for Sg. Tapai and Pantai Lido, respectively.

The oyster shell widths (cm) ranged from 3.30-6.10 and 7.63-10.6 for Sg. Tapai and Pantai Lido, respectively. The oyster shell heights (cm) ranged from 1.08-2.27 and 3.13-4.76 for Sg. Tapai and Pantai Lido, respectively.

The total soft tissue wet weight and dry weight of the two populations ranged from 0.84-4.11g (mean: 2.24 g) and 0.11-0.91g (mean: 0.42 g), respectively. Therefore, the conversion factor (0.42/2.24) of 0.19 was used to convert the dry weight into wet weight

basis. The water contents of the two populations ranged from 77.0-87.5% (mean: 81.3%).

The oysters were dissected from the shells, and they were pooled into muscle, mantle, gill and remainder. The samples were dried at 60°C until constant dry weights. They were digested in concentrated HNO₃ (Analar grade, BDH 69%). They were placed in a hot-block digester for 1 hour (40°C) and increased to 140°C for at least 3 hours (Yap *et al.*, 2003a). After dilution and filtration, they were determined

Table 1: Sampling details and some water parameters for all sampling sites.

No	Sampling Site (GPS)	Description	Temp	Cond	Sal	DO
1	Sg. Tapai (6°12'45.76"N, 102°08'38.02"E)	Recreational beach and agricultural area	30.1 ± 0.001	40419 ± 2.39	23.2 ± 0.001	1.20 ± 0.001
2	Pantai Lido (Sg. Danga) (N 01° 28.001' N; 103° 43.618' E)	Urban and agricultural areas	30.5 ± 0.001	30815 ± 4.01	17.0 ± 0.001	2.14 ±0.001

Note: Temp (°C) = Temperature; Cond= Conductivity (μ s/cm); DO= Dissolved oxygen (mg/L).

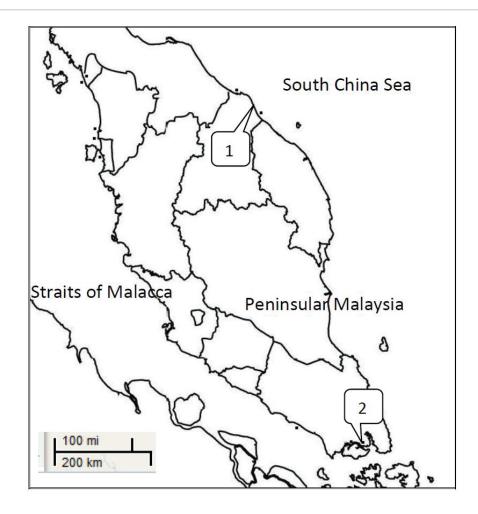


Figure 1: Map showing the sampling sites for Saccostrea cucullata in Peninsular Malaysia (1 = Sg. Tapai; 2= Pantai Lido).



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for Cd, CU, Fe, Ni, Pb and Zn by using an air-acetylene flame Atomic Absorption Spectrophotometer (AAS) Perkin-Elmer Model Analyst 800.

For quality control and quality assurance, all glassware and equipment used were acid-washed. Besides, procedural blanks and quality control samples made from standard solutions for all the six metals were analyzed to check for sample accuracy. The metal percentages of recoveries for the metals were between 80-110%. The analytical procedures for the samples were also checked with the Certified Reference Material (CRM) dogfish liver (DOLT-3, National Research Council Canada). The recoveries for the CRM were satisfactory being between 89-103%.

Health Risk Assessment

To assess a once-or long-term potential hazardous exposure to the six heavy metals through consumption of oysters (USEPA, 1989), the Estimated Daily Intake (EDI) and THQ values were calculated by using the following formulas: EDI= (Mc x consumption rate)/body weight

THQ= EDI/RfD

Where Mc is the metal concentration (mg/kg) in the oyster soft tissue (converted wet weight); body weight for Malaysian adult was 62 kg and consumption rate as 41g/day for Malaysian adults (Nurul Izzah *et al.*, 2016). The metal concentrations in μ g/g dry weight were converted into wet weight basis by using a conversion factor of 0.19 for the four different edible soft tissues of oysters.

The oral reference dose (RFD) was used to compare with the EDI (µg/kg wet weight/day) of metals in the oysters. The RFD (µg/kg wet weight/day) values used in this study were Cd: 1.00; Cu: 40.0; Fe: 700, Ni: 20 and Zn: 300, provided by the EPA's Integrated Risk Information System online database (IRIS) (IRIS, 2014). Since RFD for Pb was not available according to the EPA's IRIS (IRIS, 2014). The present study employed the RFD as 3.50 µg/kg wet weight/day, as suggested by a former study by USEPA (2000, 2008).

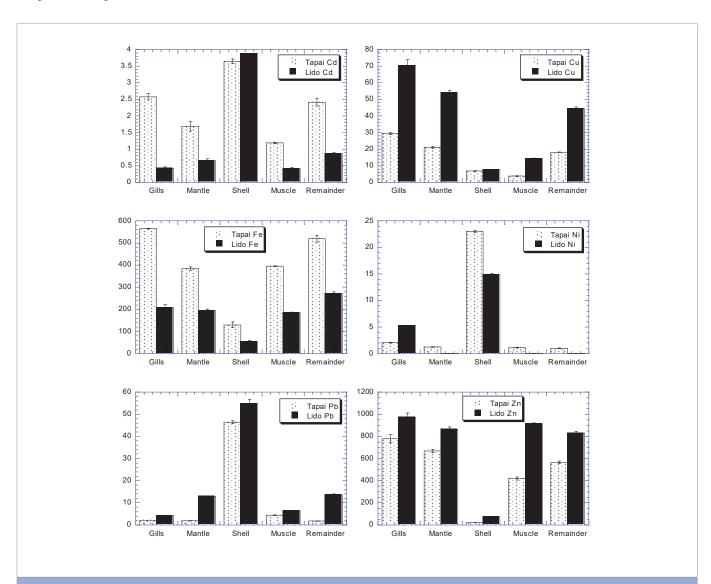


Figure 2: Heavy metal concentrations (mean ± SE, µg/g dw) in the different parts of oyster collected from Pantai Lido and Sg. Tapai.



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able 2: Values of Estimated Daily Intake (EDI) and Target Hazard Index (THQ) in the oysters collected from Sq. Tapai and Pantai Lido	

	Cd		Cu		Fe		Ni		Pb		Zn	
Tapai	EDI	THQ	EDI	THQ	EDI	THQ	EDI	THQ	EDI	THQ	EDI	THQ
Gills	0.32	0.32	3.71	0.09	70.9	0.10	0.26	0.013	0.27	0.08	98.0	0.33
Mantle	0.21	0.21	2.65	0.07	48.4	0.07	0.16	0.008	0.24	0.07	83.9	0.28
Muscle	0.15	0.15	0.48	0.01	49.8	0.07	0.14	0.007	0.55	0.16	52.8	0.18
Remainder	0.30	0.30	2.29	0.06	65.2	0.09	0.13	0.006	0.22	0.06	71.1	0.24
Lido	EDI	THQ	EDI	THQ	EDI	THQ	EDI	THQ	EDI	THQ	EDI	THQ
Gills	0.05	0.05	8.85	0.22	26.1	0.04	0.66	0.033	0.53	0.15	122	0.41
Mantle	0.08	0.08	6.78	0.17	24.4	0.03	0.01	0.001	1.63	0.47	109	0.36
Muscle	0.05	0.05	1.81	0.05	23.1	0.03	0.01	0.001	0.81	0.23	115	0.38
Remainder	0.11	0.11	5.58	0.14	34.2	0.05	0.01	0.001	1.71	0.49	104	0.35

Note: The metal concentrations in mg/kg dry weight were converted into wet weight basis by using a conversion factor of 0.19 for the four different soft tissues.

Results and Discussion

The heavy metal concentrations in the shells and four different edible soft tissues of the two oyster populations are presented in Figure 2. Based on the edible soft tissues (gills, mantle, muscle and remainder), the metal concentrations (µg/g dw) ranged from 0.42-2.58 for Cd, 3.81-70.4 for Cu, 184-565 for Fe, 0.08-5.28 for Ni, 1.73-13.6 for Pb, and 420-975.

Except for Pb, gills were found to record the highest Cd level among all the soft tissues. The remainder was found to have the highest Pb level. When compared to other soft tissues, shells recorded the highest levels of Cd, Ni and Pb, while the lowest levels of Fe and Zn in the shells.

In general, oyster accumulated the highest levels of Zn in comparison to Cd, Cu, Fe, Ni, and Pb. This is supported by the study reported for S. cucullata collected from Penang coastal waters (Yap et al., 2010). Abhilash et al. (2013) assessed the levels of Zn, Cu, Fe, Pb and Ni in S. cucullata, around Port Blair, India. Zinc was found to be the highest concentrated trace metal in all the five stations, probably due to the role of several Zn metalloenzymes in the oyster shell mineralization (Abhilash et al., 2013).

Zn is believed to be accumulated from the solute phase proportionally to the ambient Zn concentrations (Chong and Wang, 2000). This could be due to its high assimilation efficiency, and low rate constants of loss (Luoma and Rainbow, 2005).

According to Szefer et al. (2007), oysters are known to be exceptional accumulators of Zn and Cu. Similar elevated Zn levels in the oysters can be found in the literature for *S. cucullata* (Blackmore, 2001), pearl oysters (Pinctada radiata) (Gokoglu et al. (2006), and flat-tree oyster I. alatus (Saed et al., 2004).

In general, higher concentrations of Cu, Pb and Zn were found in the Pantai Lido population than Sg. Tapai population (Figure 2). Meanwhile, generally higher concentrations of Cd, Fe and Ni were found in the different parts of oyster from Sg. Tapai than Pantai Lido population. This indicated that Pantai Lido had higher bio availabilities of Cu, Pb and Zn than Sg. Tapai while Sg.

Tapai had higher bio availabilities of Cd, Fe and Ni than Pantai Lido population. By referring to the sampling site description, Pantai Lido is an urban, aquaculture area besides being a boat jetty and potentially receiving municipal wastes (Yap et al., 2006), while, Sg. Tapai is a recreational beach and agricultural area. Heavy metal bio availabilities found in these locations could be due to the anthropogenic activities found in the sites.

From the present study, the metal concentrations (µg/g dry weight) in the oyster muscles from the two populations ranged from 420-912 for Zn, 3.81-14.4 for Cu, and 4.39-6.42 for Pb. Chakraborty and Mitra A (2017) reported the metal concentrations (µg/g dry weight) muscles of S. cucullata collected from Sagar Island, West Bengal, ranging from 187-217 for Zn, 102-133 for Cu, and 13.1-17.9 for Pb. Lee et al.

(2015) suggested that the adductor muscle of Atrina japonica displayed an essential role in the energy loading. Uddin et al. (2007) reported that the weight loss of the muscle of scallop during the spawning was related to energy storage. Baik et al. (2001) reported that the adductor muscle of A. japonica had higher glycogen content than those in the visceral mass.

Values of EDI and THQ for the six metals in the two populations are given in Table 2. Overall, the values of EDI ranged from 00.05-0.32 for Cd, 0.48-8.85 for Cu, 23.1-70.9 for Fe, 0.01-0.66 for Ni, 0.22-1.71 for Pb, and 52.8-122 for Zn. The values of THQ for all the six metals are all below 1.0.

This means that the daily consumption of oysters collected from Sg. Tapai and Pantai Lido would not likely result in adverse health effects during a lifetime in a human population (Bogdanovic et al., 2014). This also indicated that the two oyster populations from the six sites are safe, with no non-carcinogenic effects of the six metals, for consumption at least based the sampling period.

Conclusion

This study determined the concentrations of six heavy metals in the shells and four edible soft tissues of S. cucullata collected from Sg. Tapai and Pantai Lido obtained in 2008. Higher bio availabilities of Cu, Pb and Zn to S. cucullata were found in the Pantai Lido population; meanwhile, Sg. Tapai had higher bio availabilities of Cd, Fe and Ni to the oyster population.

For HHRA, the THQ values of all the six metals are below than 1.0. These values indicated that the edible soft tissues of oysters would cause no non-carcinogenic risk of Cd, Cu, Fe, Ni, Pb and Zn to the



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consumers. Therefore, rock oysters are safe to be consumed based on the samples collected in 2008.

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